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Habitat complexity influenced mixed-species bird flock composition and occurrence in highlands

Pablo Muñoz^{1,2};  <https://orcid.org/0000-0002-3030-1010>

Luis Sandoval^{1,3*};  <https://orcid.org/0000-0002-0793-6747>

1. Laboratorio de Ecología Urbana y Comunicación Animal, Escuela de Biología, Universidad de Costa Rica, San José, Costa Rica; biosandoval@hotmail.com (*Correspondencia); pablomgl94@gmail.com
2. Organización para Estudios Tropicales. Oficina de Costa Rica, Universidad de Costa Rica, San José, Costa Rica.
3. Colección de Ornitología, Museo de Zoología, CIBET, Universidad de Costa Rica, San José, Costa Rica.

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ABSTRACT

Introduction: Mixed-species bird flocks are groups that change rapidly throughout the day, influenced by factors such as resource availability, vocalizations, and habitat complexity. Habitat complexity can either enhance or limit interspecific interactions by increasing the number of species or individuals that interact. However, variations in habitat structure are often overlooked when studying mixed-species flock composition, leading to incomplete or unrealistic interpretations of factors influencing flock composition.

Objective: This study aims to investigate the relationship between the composition and occurrence of mixed-species bird flocks and the habitat complexity of a highland forest in Costa Rica.

Methods: We conducted this study along a 5.5 km transect in the highland area of Braulio Carrillo National Park in Heredia, Costa Rica (10°05' N & 84°04' W; 2 100 m a.s.l.) from March to December 2019. Mixed-species flocks were surveyed twice weekly from 06:00 to 11:00 h. Habitat complexity was assessed using 16 plots (10x10 m) distributed across the study area to capture most of the habitat variation. We estimated the probability of occurrence at different habitat types for mixed-species flock that vary in sizes using kernel density estimation in QGIS.

Results: A total of 34 species were recorded across 125 mixed-species flocks, categorized into 50 small-sized, 46 medium-sized, and 30 large-sized flocks. Flock sizes were associated with different habitat characteristics. A positive association was found between higher canopy and understory cover, tree quantity, and diameter at breast height (DBH) with species composition and abundance in large-sized flocks, but not in medium- and small-sized flocks.

Conclusion: While the overall presence of mixed-species flocks was not limited by habitat structure, their size and composition were significantly associated with habitat.

Key words: endemic birds; highland birds; oak forest; peat bogs; tanagers.

RESUMEN

La complejidad del hábitat influye la composición y la ocurrencia de bandadas mixtas de aves en las tierras altas

Introducción: Las bandadas de aves mixtas son grupos que cambian rápidamente a lo largo del día, influenciados por factores como la disponibilidad de recursos, vocalizaciones y complejidad del hábitat. La complejidad del hábitat puede favorecer o limitar las interacciones interespecíficas al aumentar el número de especies o individuos que interactúan. Sin embargo, las variaciones en la estructura del hábitat a menudo no se incluyen cuando



se estudia la composición de las bandadas de especies mixtas, lo que lleva a interpretaciones incompletas de los factores que influyen en la composición de la bandada.

Objetivo: Este estudio tiene como objetivo investigar la relación entre la composición y la ocurrencia de bandadas mixtas de aves y la complejidad del hábitat de un bosque de tierras altas en Costa Rica.

Métodos: Realizamos este estudio a lo largo de un transecto de 5,5 km en el Parque Nacional Braulio Carrillo en Heredia, Costa Rica (10°05' N & 84°04' O; 2 100 m s. n. m.) de marzo a diciembre de 2019. Las bandadas de especies mixtas se muestrearon dos veces por semana de 06:00 a 11:00 h. La complejidad del hábitat se evaluó utilizando 16 parcelas (10x10 m) distribuidas en el área de estudio para capturar la mayor variación de los hábitat. Estimamos la probabilidad de ocurrencia en diferentes tipos de hábitat para bandadas de especies mixtas que varían en tamaño utilizando la estimación de densidad kernel en QGIS.

Resultados: Se registraron un total de 34 especies en 125 bandadas de especies mixtas, categorizadas en 50 bandadas de tamaño pequeño, 46 de tamaño mediano y 30 de tamaño grande. Los tamaños de las bandadas se asociaron con diferentes características del hábitat. Se encontró una asociación positiva entre una mayor cobertura de dosel y sotobosque, cantidad de árboles y diámetro a la altura del pecho (DAP) con la composición y abundancia de especies en bandadas de tamaño grande, pero no en bandadas de tamaño mediano y pequeño.

Conclusión: Si bien la presencia general de bandadas de especies mixtas no estuvo limitada por la estructura del hábitat, su tamaño y composición se asociaron significativamente con el hábitat.

Palabras clave: aves endémicas; aves de las tierras altas; bosques de robles; turberas; tangaras.

INTRODUCTION

Habitat complexity, defined as the physical structure of the habitat (Carter et al., 2018), significantly influences species assembly and their interactions within a community, ranging from parasite-host dynamics to food webs and predator-prey interactions (Carter et al., 2018; Kovalenko et al., 2012; Staniczenko et al., 2017). Species interactions may be enhanced in habitats with high complexity, as the abundant microhabitats and resources support a greater number and diversity of species (Maldonado-Coelho & Marini, 2000; Maldonado-Coelho & Marini, 2004). Community interactions are often assessed under the assumption of homogeneity in individual and habitat structure, overlooking their variations and leading to incomplete insights or unrealistic interpretations (Carter et al., 2018). Therefore, it is crucial to consider this variation, as heterogeneity is essential for understanding the drivers of complex communities (Carter et al., 2018; Gibert & Brassil, 2014; Mokross et al., 2018). This is particularly relevant when studying bird communities, which not only utilize habitat structures differently but also tend to form complex interspecies associations, such as mixed-species flocks within these habitats (Jones & Robinson, 2021).

Mixed-species flocks of birds are among the most common and interactive group systems in nature (Goodale et al., 2020; Morse, 1970). These flocks are dynamic assemblages with both positive and negative interactions (Goodale et al., 2020). Understanding how they are established, maintain cohesion, and interact intra-specifically (Jones et al., 2020) is central to community ecology. The most commonly reported interactions within mixed-species flocks include behavioral interactions (Morse, 1970), role turnovers (Farine, 2021), and species facilitation in joining the flock (Martínez et al., 2018; Muñoz 2022). Role turnovers within the flock occur when the relative population of a species increases in the area, like *Atlapetes tibialis* that becomes a gregarious species outside the breeding season and assumes a leadership role in mixed-species flocks (Muñoz 2022). Facilitation occurs when members of the flock display behaviors that trigger flocking, such as specific vocalizations (Pagani-Núñez et al., 2018), and can also result from external factors like habitat complexity (Croft et al., 2011; Goodale et al., 2010).

Habitat complexity may facilitate flocking behavior in various ways. For instance, specific habitat traits (e.g., background conspicuousness) influence signal transmission and its

reception by conspecifics and heterospecifics (Munn, 1985; Uy & Stein, 2007). Moreover, habitat structural traits, such as forest height and area size, influence the gregariousness of mixed-species flocks (Mokross et al., 2014), as they can physically limit the number and identity of species joining a flock and affect the cost-benefit tradeoff for potential flock members. For example, the absence of larger trees limits the occurrence of woodcreepers, which used them to forage or establish a territory (Darrah & Smith, 2013; Maldonado-Coelho & Marini, 2000), while the lack of dense understory restricts wrens and antbirds that rarely move into open areas (Mokross et al., 2018). Other aspects of forest structure that might affect mixed-species flock composition, such as resource availability (e.g., fruits or flowers), shelter (e.g., abundance and size of trees), or specialized foraging structures (e.g., dense understory, canopy cover), have not been fully considered (Hutto, 1988; Jones, & Robinson, 2021; Kotagama & Goodale, 2004; Mangini et al., 2023b). Changes in these habitat traits may trigger composition effects on mixed-species flock structure. Reduced availability of understory resources may decrease the complexity of mixed-species flocks by reducing the overall foraging efficiency of bird species (Hutto, 1988). Additionally, a complex habitat structure with a higher canopy and cover may provide more usable habitat space for bird species when

moving in a mixed-species flock (Mokross et al., 2014). Changes in habitat complexity, such as higher or lower understory cover, can also limit the species that can potentially join the mixed-species flock due to reduction of forest cover for them (Rutt et al., 2020).

Habitat complexity varies significantly not only at a large scale, but also over short distances within the same area (Rutt et al., 2020) and this will affect directly the composition and level of association of mixed-species flocks. Our objective is therefore to describe the correlation between the composition and occurrence sites of mixed-species flocks with the habitat complexity (structure) of a highland forest. Costa Rican highlands are suitable to answer this question because they present mixed-species flocks all year round with its members interacting among them at all times (Powell, 1979; Powell, 1985). In addition, tropical mixed-species flocks tend to use all forest structures present in its habitat, like branches, trunks, mosses, understory cover, and treetops (Powell, 1985).

MATERIALS AND METHODS

We conducted this study along a 5.5 km transect (Fig. 1) in the highland area of Braulio Carrillo National Park in Heredia, Costa Rica (10°05'N & 84°04'W; 2100 m a.s.l.) from March to December 2019. The study area includes

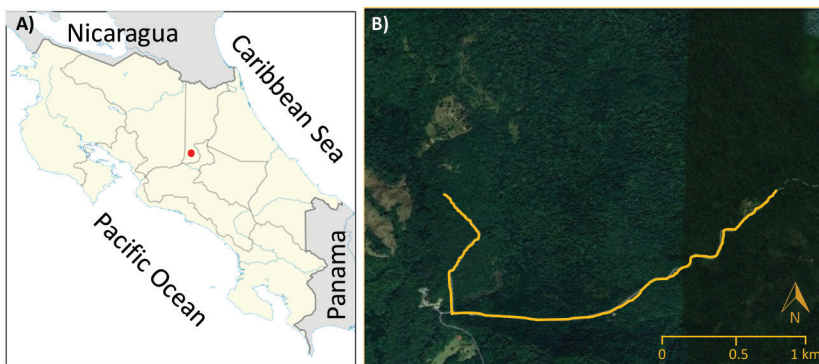


Fig. 1. A. The red dot is the study location area in Costa Rica. B. Study area map with the transect used for sampling highlighted in white.



a mix of oak (*Quercus* sp.) and Andean alder (*Alnus acuminata*) forest, secondary growth areas (dominated by bamboos, Malvaceae's trees and bushes, and Ericaceae bushes), and several disturbed areas featuring pastures, peat bogs, and isolated trees. These features create a heterogeneous landscape that allowed comparisons between mixed-species flocks structure according to the habitat composition.

We searched for mixed-species flocks twice per week from 06:00 to 11:00 h, along the transect (Fig. 1). To minimize potential biases related to time and location, we alternated the starting point each day, beginning each sampling session at the opposite end of the transect from the previous one. Whenever we encountered a flock, we followed for 10 min and recorded the maximum number of individuals per species during this time period. The geographical position of each flock was recorded at the beginning and end of this period using a Garmin 64s GPS device.

We separated the mixed-species flocks in three groups based on the number of individuals and species (Chen & Hsieh, 2002; Mangini et al., 2023a). Small-size mixed-species flocks included <4 species or <8 individuals. Medium-size mixed-species flocks included 5-6 species or 9-11 individuals. Large-size mixed-species flocks included >7 species or >12 individuals. This information allowed us to compare if size of mixed-species flocks size varied according to habitat structure and characteristics.

Relationship between flocks and habitats: To measure habitat complexity traits, we used 16 plots of 10x10 m distributed across the study area to cover most common habitat types present (Fig. 1). Half of the plots were located near the transect (<50 m from the transect edge) where mixed-species flocks were sampled, and the other half of the plots were located at least 200 m away from the transects. We placed the plots ensuring a minimum distance of 200 m between them and to have four replicates in each habitat types present in the study area: pastures/peat bogs, old-growth forest, forest plantation, and secondary forest. In

each plot, we counted all trees with the diameter at breast height (DBH) greater than 10 cm and measured the exact DBH for each one. We recorded canopy height at three points within the plot (one at the center and two sides) using a laser rangefinder. Additionally, we selected four equidistant 1x1 m sub-plots within each plot and estimated understory coverage as a percentage of ground cover by dead leaves, branches, and plants, but not grasses. Finally, we measured the leaf area index (LAI) using five photos per plot, following the method published by Martin (2015). The photos were taken with a Canon SL1 camera and an 18-55 mm lens: four photos directed at each cardinal point at a 45-degree angle and one photo at a 90-degree angle towards the canopy. For the eight plots near the trail, we placed 1x1 m sub-plots to measure understory cover at each of the four corners of the 10x10 m plots. For the eight plots farther from the trail, we positioned the 1x1 m sub-plots along the sides of the 10x10 m plots, midway between each corner. Using the data gathered from all 16 plots, we interpolated habitat structure traits to the rest of the un-sampled area using the inverse distance weighting method in QGIS (QGIS Development Team, 2021). This created a separate layer for each habitat structure trait measured, covering the entire study area (Fig. 2), and allowed us to: 1) determine the suitability of different habitat structures in the study area for each mixed-species flock, and 2) assign the habitat structure where each observed flock occurred.

Before analyzing the mixed-species flock occurrence, we ensured that each flock included in the analysis were at least 10 m apart from other flock. We used this distance because the predictor layers employed were interpolated to 10x10 m pixels. We estimated the probability of occurrences of each mixed-species flock size group in the study area using kernel density estimation in QGIS (QGIS Development Team, 2021). For this, we mapped the occurrences of each mixed-species flock size on top the five cover layers of habitat complexity (see above), and created three probability map models, one for each mixed-species flock size. Finally, we

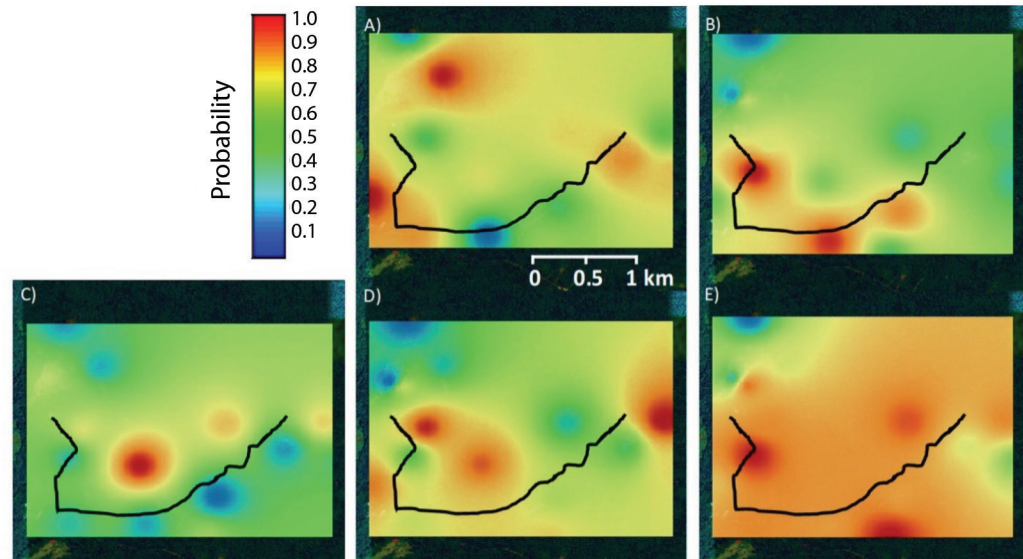


Fig. 2. Inverse weighted interpolations of habitat structure variables for the whole area; **A.** Understory cover, **B.** Tree height, **C.** Diameter at breast height, **D.** Number of trees, and **E)** Leaf area index. Red indicates higher values of the variable, while the blue indicates lower values of the variable.

used the GPS position of each mixed-species flock to quantify its occurrence in each main habitat of the study area: pasture/peat bogs, old-growth, forest plantation, and secondary forest; according to habitat observed in Google maps on January 2020, and in cases where habitat was covered by clouds we assigned the category of unknown habitat.

Statistical analysis: First, we conducted three chi-square test to compare if each mixed-species flock group-size used different the main habitats (i.e., pasture/peatbogs, old-growth, forest plantation, and secondary forest). Second, we compared if the habitat structure where each mixed-species flock group-size occurred varied using a multiple analysis of variance (MANOVA). For this analysis, we used as our independent variable the flock size category. We used as dependent variables the values of each of the five habitat structure measurements (i.e., DBH, canopy height, tree abundance, understory coverage, and LAI) where each flock occurred. This value was estimated using the layer for each habitat structure trait built using the inverse

distance weighting method in QGIS. For this analysis, we used a posterior comparison to report which habitat structure variable were different between mixed-species flock group-size. Third, we conducted a non-metric multidimensional scaling (NMDS) analysis with Euclidian distance and 10 000 permutations to compare if the mixed-species flock groups varied in composition (i.e., species and abundance) and its relationship with the five habitat structure variables. Finally, we used a one-way PERMANOVA to analyze if the three mixed-species flock groups were significantly different in the NMDS, and with pair-wise comparison we analyzed differences between flock sizes. All statistical analysis were conducted using PAST 4.17 (<https://www.nhm.uio.no/english/research/resources/past/>).

RESULTS

We recorded 34 species in 127 mixed-species flocks, with 51 classified as small-size, 46 as medium-size, and 30 as large-size (Table 1, Appendix 1). We found that each mixed-species



Table 1
Percentage of occurrence of mixed-species flock size in different habitat types.

Habitat Type	Small (%)	Medium (%)	Large (%)
Pasture/Peatbogs	20.8	15.7	34
Old-growth	23.1	27.8	14.9
Forest Plantation	19.8	32.5	40.4
Secondary Forest	15.4	7.2	2.1
Unknown	20.8	16.9	8.5

flock size used differently the four main habitats in the study area. Small-size mixed-species flocks used in similar frequency the four main habitats of the study site ($X^2 = 0.80$, $df = 3$, $P = 0.85$; Table 1). Medium-sized mixed-species flocks used more old-growth and forest plantations than the other two main habitats ($X^2 = 8.84$, $df = 3$, $P = 0.03$; Table 1). Meanwhile, large-size mixed-species flocks used more forest plantations and pasture/peat bog habitats than the other two main habitats ($X^2 = 12.30$, $df = 3$, $P = 0.006$; Table 1).

Each mixed-species flock group sizes occurred in sites with different habitat structure according to the five measurements took (MANOVA: Wilk's $\lambda = 0.92$, $F_{8,438} = 2.26$, $P = 0.02$; Fig. 3). Among the habitat structure variables analyzed independently with the posterior comparisons of the MANOVA, only the diameter at breast height (DBH) showed a significant difference between mixed-species flock group sizes ($F_{2,222} = 3.70$, $P = 0.03$). Large-size mixed-species flocks were more often found in sites with larger DBH compared to the other two mixed-species flock group sizes (Fig. 4). Meanwhile, the mixed-species flock group sizes used sites with similar values for the other four habitat structure variables ($F_{2,222} < 0.82$, $P > 0.44$ for all variables; Fig. 4).

We found a positive association between higher canopy and understory cover, tree quantity, and DBH with species composition and abundance per species in large-size mixed-species flocks (PERMANOVA: $F = 4.60$, $P < 0.001$; Fig. 4). However, small and medium-size

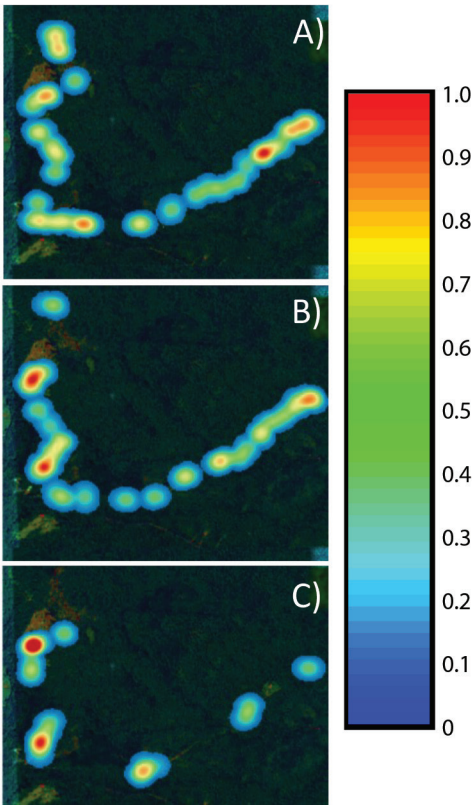


Fig. 3. Kernel density estimation models for A. Small-sized, B. Medium-sized, and C. Large-sized mixed-species flocks. Red indicates higher probabilities that each flock size occur in that site, while blue indicates lower probabilities of the occurrence.

mixed-species flocks were associated with sites that showed higher values of leaf area index (LAI; Fig. 4). Additionally, small and medium-size mixed-species flocks had similar species composition and abundance per species (pair-wise comparison: $P = 0.13$), but both differed significantly from large-size mixed-species flocks (small vs. large: $P < 0.001$, medium vs. large: $P < 0.001$; Fig. 4).

DISCUSSION

Habitat structure affects the size of flocks that occur in different sites of the study area. This pattern was expected, since more complex

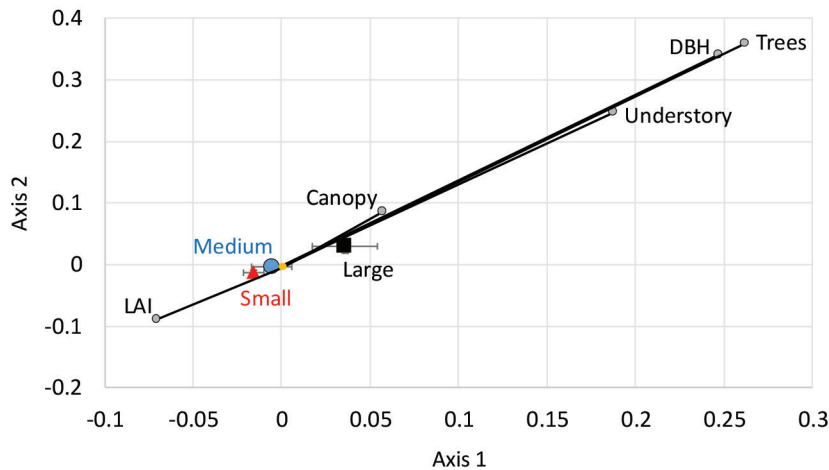


Fig. 4. Non-Metric Multidimensional Scaling (NMDS) plot showing the position of different mixed-species flocks sizes and its relationship with the habitat complexity variables. Error bars represent one standar error in both axes.

habitat structures facilitate larger mixed-species flocks by accommodating species with varying microhabitat preferences, such as understory specialists or foraging specialists (Zhang et al., 2013; Zuluaga & Rodewald, 2015). For instance, woodcreepers require large trees with mosses and bromeliads, commonly found in old secondary forests and plantations in the study area, to join the flocks (Cintra et al., 2006; Rodrigues et al., 2016). Additionally, wrens need dense understories to establish territories and join flocks as they pass through (Caro et al., 2013; Shogren et al., 2019). Consequently, habitats with a greater variety of complex structures (e.g., higher canopy and understory coverage, and more trees) allowed more species to join the flocks. This explain why we observed more medium and large mixed-species flocks in areas with forest plantations and old-growth forests (Zuluaga & Rodewald, 2015). However, large mixed-species flocks were also found in higher numbers in open disturbed areas (i.e., pastures/peat bogs), which have less complex habitat characteristics (Harris et al., 2018; Lindgren et al., 2018). This fiding can be explained by the fact that large flocks in open areas increased in size by adding more individuals of the same species, rather than different species as seen in forest habitats (pers. obs.). This occurs because

species that inhabit in open areas (i.e., mixed-species core species) and form mixed-species flocks are typically more numerous such as chlorospingus and warblers (Blake & Loiselle, 2001). Contrary to previous reports from another tropical forest, where flock complexity (i.e., more individuals and species) increased on advanced forest successional stages (Zhang et al., 2013). Consequently, our highland mixed-species flock system showed that large-size flocks may occur on both old forest plantations and peat bogs

Conversely, small mixed-species flocks occurred in similar abundance across all habitat types. This is because the core species that form Costa Rican highland mixed-species flocks (e.g., *Tangara dowii*, *Chlorospingus pileatus*, *Margarornis rubiginosus*, and *Atlapetes tibialis*, Appendix 1) are present in all habitats, including open areas, forest edges, secondary forests, and plantations (Barrantes, 2009; Barrantes et al., 2011; Stiles & Skutch, 1989). Additionally, species from genera that did not usually involve in mixed-species flocks in lowlands as *Turdus*, *Piranga* and *Atlapetes* that are common in open areas and secondary forest canopies also join mixed-species flocks (Bohórquez, 2003, Powell, 1979, Stiles & Skutch, 1989), probably because increase the food intake and decrease



the probabilities of predation (Hino, 2000; Mangini et al., 2023b). Consequently, mixed-species flocks occurrences are not limited by habitat structure in highlands (Barrantes et al., 2011) and occurred in all the sites, allowing small-size flocks use all habitats available in the study area.

When comparing habitat complexity and its relationship with flock composition, we found that medium and large-sized flocks did not respond uniformly and used microhabitat inside the four habitat types of the study area differently. While, medium-sized flocks were more similar in structure and habitat use to small-sized flocks. This similarity may be due to both flock sizes sharing more species and having similar numbers of individuals per species (Appendix 1), especially if both are primarily composed of core flock species that tend to forage on substrates such as isolated trees, forest edges, and mature forests (Barrantes & Pereira, 2002; Stiles & Skutch, 1989). Abrupt changes in habitat structure over short distances (e.g., transitioning from old-growth forest or forest plantation to open areas), as observed in the study area (Fig. 1), affected the number of species in our mixed-species flocks but not the number of individuals. Forest-dependent species (e.g., woodcreepers, tapaculos, wrens) would abandon the mixed-species flocks upon reaching open areas (Rutt et al., 2020), while core species continued foraging together across different habitats. Consequently, the number and type of species that form small and medium-size flocks were more similar to each other than to larger flocks. In other environments, such as the Amazonian lowlands, abrupt changes in habitat structure generally limit the movement of mixed-species flocks because their core species are more forest-dependent than what we found for Costa Rican highland species (Rutt et al., 2020). Therefore, on those sites mixed-species flocks composition in each habitat type are different, showing a greater association between the species that form the

mixed-species flocks and the habitats in which they occur (Zhang et al., 2013).

We concluded that the presence of mixed-species flocks as a whole was not limited by forest structures in the study area, but their sizes and composition were associated with habitat complexity. Forest structures with more microhabitats for species to forage may enhance the occurrence of large-size mixed-species flocks in an area. A core set of species that joined mixed-species flocks are consistent across all forest types and flock sizes, which may explain why these flocks were present throughout the study area. Forest structure likely might also affect the number and type of interactions of a flock members, but this is an aspect that remain understudied (Zhang et al., 2013).

Ethical statement: the authors declare that they all agree with this publication and made significant contributions; that there is no conflict of interest of any kind; and that we followed all pertinent ethical and legal procedures and requirements. All financial sources are fully and clearly stated in the acknowledgments section. A signed document has been filed in the journal archives.

See supplementary material
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REFERENCES

- Barrantes, G. (2009). The role of historical and local factors in determining species composition of the highland avifauna of Costa Rica and western Panamá. *Revista de Biología Tropical*, 57, 333–346.
- Barrantes, G., & Pereira, A. (2002). Seed dissemination by frugivorous birds from forest fragments to adjacent pastures on the western slope of Volcán Barva, Costa Rica. *Revista de Biología Tropical*, 50(2), 569–575.
- Barrantes, G., Yglesias, M., & Fuchs, E. J. (2011). The roles of history and habitat area in the distribution and composition of avian species assemblages in the highlands of Costa Rica. *Journal of Tropical Ecology*, 27, 421–428.
- Blake, J. G., & Loiselle, B. A. (2001). Bird assemblages in second-growth and old-growth forests, Costa Rica: perspectives from mist nets and point counts. *Auk*, 118, 304–326.
- Bohórquez, C. I. (2003). Mixed-species bird flocks in a montane cloud forest of Colombia. *Ornitología Neotropical*, 14, 67–78.
- Caro, L. M., Caycedo-Rosales, P. C., Bowie, R. C. K., Slabbekoorn, H., & Cadena, C. D. (2013). Ecological speciation along an elevational gradient in a tropical passerine bird?. *Journal of Evolutionary Biology*, 26, 357–374.
- Carter, S. K., Vodopich, D., & Crumrine, P. W. (2018). Heterogeneity in body size and habitat complexity influence community structure. *Journal of Freshwater Ecology*, 33, 239–249.
- Chen, C. C., & Hsieh, F. (2002). Composition and foraging behaviour of mixed-species flocks led by the Grey-cheeked Fulvetta in Fushan Experimental Forest, Taiwan. *Ibis*, 144, 317–330.
- Cintra, R., Maruoka, A. E., & Naka, L. N. (2006). Abundance of two Dendrocincla woodcreepers (Aves: Dendrocolaptidae) in relation to forest structure in Central Amazonia. *Acta Amazonica*, 36, 209–219.
- Croft, D. P., Madden, J. R., Franks, D. W., & James, R. (2011). Hypothesis testing in animal social networks. *Trends in Ecology & Evolution*, 26, 502–507.
- Darrah, A. J., & Smith, K. G. (2013). Comparison of foraging behaviors and movement patterns of the Wedge-billed Woodcreeper (*Glyphorhynchus spirurus*) traveling alone and in mixed-species flocks in Amazonian Ecuador. *Auk*, 130(4), 629–636.
- Farine, D. R. (2021). Structural trade-offs can predict rewiring in shrinking social networks. *Journal of Animal Ecology*, 90, 120–130.
- Gibert, J. P., & Brassil, C. E. (2014). Individual phenotypic variation reduces interaction strengths in a consumer–resource system. *Ecology and Evolution*, 4, 3703–3713.
- Goodale, E., Beauchamp, G., Magrath, R. D., Nieh, J. C., & Ruxton, G. D. (2010). Interspecific information transfer influences animal community structure. *Trends in Ecology & Evolution*, 25, 354–361.
- Goodale, E., Sridhar, H., Sieving, K. E., Bangal, P., Colorado Z, G. J., Farine, D. R., Heymann, E. W., Jones, H., Krams, I., Martínez, A. E., Montañó-Centellas, F., Muñoz, J., Srinivasan, U., Theo, A., & Shanker, K. (2020). Mixed company: a framework for understanding the composition and organization of mixed-species animal groups. *Biological Reviews*, 95, 889–910.
- Harris, V., Kendal, D., Hahs, A. K., & Threlfall, C. G. (2018). Green space context and vegetation complexity shape people's preferences for urban public parks and residential gardens. *Landscape Research*, 43, 150–162.
- Hino, T. (2000). Intraspecific differences in benefits from feeding in mixed-species flocks. *Journal of Avian Biology*, 31, 441–446.
- Hutto, R. L. (1988). Foraging behavior patterns suggest a possible cost associated with participation in mixed-species bird flocks. *Oikos*, 51, 79–83.
- Jones, H. H., & Robinson, S. K. (2021). Vegetation structure drives mixed-species flock interaction strength and nuclear species roles. *Behavioral Ecology*, 32, 69–81.
- Jones, H. H., Walters, M. J., & Robinson, S. K. (2020). Do similar foragers flock together? Nonbreeding foraging behavior and its impact on mixed-species flocking associations in a subtropical region. *Auk*, 137, 1–16.
- Kotagama, S. W., & Goodale, E. (2004). The composition and spatial organization of mixed-species flocks in a Sri Lankan rainforest. *Forktail*, 20, 63–70.
- Kovalenko, K. E., Thomaz, S. M., & Warfe, D. M. (2012). Habitat complexity: approaches and future directions. *Hydrobiologia*, 685, 1–17.
- Lindgren, J., Kimberley, A., & Cousins, S. A. (2018). The complexity of forest borders determines the understory vegetation. *Applied Vegetation Science*, 21, 85–93.
- Mangini, G. G., Gandoy, F. A., Areta, J. I., & Blendinger, P. G. (2023a). Benefits of foraging in mixed-species flocks depend on species role and foraging strategy. *Ibis*, 165, 629–646.
- Mangini, G. G., Rutt, C. L., Sridhar, H., Buitron, G., Muñoz, J., Robinson, S. K., Montañó-Centellas, F., Zarco, A., Fanjul, M. E., Fernandez-Arellano, G., Xing, S., & Camerlenghi, E. (2023b). A classification scheme for mixed-species bird flocks. *Philosophical Transactions of the Royal Society B*, 378, 20220100.
- Maldonado-Coelho, M., & Marini, M. Â. (2000). Effects of forest fragment size and successional stage on



- mixed-species bird flocks in southeastern Brazil. *Condor*, 102, 585–594.
- Maldonado-Coelho, M., & Marini, M. A. (2004). Mixed-species bird flocks from Brazilian Atlantic forest: the effects of forest fragmentation and seasonality on their size, richness and stability. *Biological Conservation*, 116, 19–26.
- Martin, C. A. (2015). *LAI: calculate indirect leaf area index (LAI) from images. R package version 0.0.0.9004*. <https://github.com/cmartin/LAI>. DOI:10.5281/zenodo.34690.
- Martínez, A. E., Pollock, H. S., Kelley, J. P., & Tarwater, C. E. (2018). Social information cascades influence the formation of mixed-species foraging aggregations of ant-following birds in the Neotropics. *Animal Behaviour*, 135, 25–35.
- Mokross, K., Potts, J. R., Rutt, C. L., & Stouffer, P. C. (2018). What can mixed-species flock movement tell us about the value of Amazonian secondary forests? Insights from spatial behavior. *Biotropica*, 50, 664–673.
- Mokross, K., Ryder, T. B., Côrtes, M. C., Wolfe, J. D., & Stouffer, P. C. (2014). Decay of interspecific avian flock networks along a disturbance gradient in Amazonia. *Proceedings of the Royal Society B: Biological Sciences*, 281, 20132599.
- Morse, D. H. (1970). Ecological aspects of some mixed-species foraging flocks of birds. *Ecological Monographs*, 40, 119–168.
- Munn, C. A. (1985). Permanent canopy and understory flocks in Amazonia: species composition and population density. *Ornithological Monographs*, 36, 683–712.
- Muñoz, P. (2022). *Patrones de co-ocurrencia, elección de hábitat y comunicación heteroespecífica dentro de bandadas mixtas de aves* (Tesis de Maestría). Universidad de Costa Rica, Sistema de Estudios de Posgrado.
- Pagani-Núñez, E., Xia, X., Beauchamp, G., He, R., Husson, J. H., Liang, D., & Goodale, E. (2018). Are vocal characteristics related to leadership patterns in mixed-species bird flocks? *Journal of Avian Biology*, 2018, e01674.
- Powell, G. V. (1979). Structure and dynamics of interspecific flocks in a Neotropical mid elevation forest. *Auk*, 96, 375–390.
- Powell, G. V. (1985). Sociobiology and adaptive significance of interspecific foraging flocks in the Neotropics. *Ornithological Monographs*, 36, 713–732.
- QGIS Development Team. (2021). *QGIS geographic information system* (Version 3.20). Open Source Geospatial Foundation Project. <https://qgis.org>
- Rodrigues, D. J., Florêncio, F. P., Oliveira, J., Oliveira, D. M., Lollback, G. W., & Hero, J. M. (2016). Habitat associations of woodcreeper (Aves: Dendrocolaptidae) assemblage in selectively logged areas of Southern Amazonia. *Journal of Tropical Ecology*, 32, 63–74.
- Rutt, C. L., Mokross, K., Kaller, M. D., & Stouffer, P. C. (2020). Experimental forest fragmentation alters Amazonian mixed-species flocks. *Biological Conservation*, 242, 108415.
- Shogren, E. H., Jones, M. A., Sandercock, B. K., & Boyle, W. A. (2019). Apparent survival of tropical birds in a wet, premontane forest in Costa Rica. *Journal of Field Ornithology*, 90, 117–127.
- Staniczenko, P. P., Lewis, O. T., Tylianakis, J. M., Albrecht, M., Coudrain, V., Klein, A. M., & Reed-Tsochas, F. (2017). Predicting the effect of habitat modification on networks of interacting species. *Nature Communications*, 8, 1–10.
- Stiles, F. G., & Skutch, A. F. (1989). A guide to the birds of Costa Rica. *Cornell University Press*, 632.
- Uy, J. A. C., & Stein, A. C. (2007). Variable visual habitats may influence the spread of colourful plumage across an avian hybrid zone. *Journal of Evolutionary Biology*, 20, 1847–1858.
- Zhang, Q., Han, R., Huang, Z., & Zou, F. (2013). Linking vegetation structure and bird organization: response of mixed-species bird flocks to forest succession in subtropical China. *Biodiversity and Conservation*, 22, 1965–1989.
- Zuluaga, G. J. C., & Rodewald, A. D. (2015). Response of mixed-species flocks to habitat alteration and deforestation in the Andes. *Biological Conservation*, 188, 72–81.